



# **Mitigation technologies and practices for reducing CH<sub>4</sub> emissions from animal manure management**

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For manure management sector, CH<sub>4</sub> was produced when the organic matter in the manure was anaerobically decomposed by the methanogens. Therefore, CH<sub>4</sub> emission can be happened during the whole manure management chain (MMC), from manure excreted, and the following manure storage and treatment sector, and even in the land application sector when the anaerobically condition provided. Based on the literature review for the CH<sub>4</sub> mitigation options used in the major animal categories, including pig, dairy, beef cattle and poultry (Table 1-4), the mitigation technologies, practices and the possible integrated mitigation technologies were summarized, and the best technology recommended for each stage was also summarized (Table 5). Meanwhile, the possible effects of the technologies on N<sub>2</sub>O and NH<sub>3</sub> emissions were also evaluated simultaneously.

## **1 CH<sub>4</sub> mitigation technologies used in each stage of the manure management chain**

### **1.1 In house sector**

The CH<sub>4</sub> mitigation options used in this sector including change manure characteristics by manipulating animal diet, changing to a better manure cleaning system, and reducing the exhausted CH<sub>4</sub> emission to the ambient environment by using biofilter.

### 1.1.1 Manipulating animal diet

Wu-Haan et al. (2007) reported that the emission reduction diet (RE diet) containing 6.9% of a CaSO<sub>4</sub>-zeolite mixture and slightly reduced crude protein (CP) can reduce the CH<sub>4</sub> emission by 10.3-36% in hen facility. Kim et al. (2004) found that an acidogenic Ca and P source (CaSO<sub>4</sub>-H<sub>3</sub>PO<sub>4</sub>) in swine diets could decrease the urinary pH and reduce CH<sub>4</sub> emission by 14% from swine facilities. CaSO<sub>4</sub> was added to the diet as an acidifying agent, which caused the decrease of pH. A reduction in manure pH or perhaps reduced manure VFA could explain the reduced CH<sub>4</sub> emission based on the RE diet (Wu-Haan et al., 2007). However, another acidifying diet by adding CaCl<sub>2</sub> caused the in-house CH<sub>4</sub> emission increased by 73% (Eriksen et al., 2014). Meanwhile, the reduced manure pH caused by the acidifier diet led to the NH<sub>3</sub> emission being reduced by 27-46% (Wu-Haan et al., 2007; Eriksen et al., 2014).

On the other hand, Wiedemann et al. (2016) reported that the CH<sub>4</sub> emission can be reduced by 26% by using low CP diet in layer production. However, Menezes et al. (2016) found there is no significant influence on manure CH<sub>4</sub> emissions by using LCP diet in beef cattle production.

### 1.1.2 Optimization for manure cleaning system

Increasing manure cleaning frequency can reduce the manure retention time in animal house, and therefore may reduce the CH<sub>4</sub> emission. Ulens et al. (2014) found the in-house CH<sub>4</sub> emission was reduced by 39% by increasing the manure cleaning frequency. Wang et al. (2017) compared the in-house CH<sub>4</sub> emission factor of four typical pig manure collection method based on meta-analysis. The results show that different in-house manure collection methods have a significant impact on gas emissions, especially for CH<sub>4</sub> and N<sub>2</sub>O. The CH<sub>4</sub> EF is largest for the deep-pit mode (median value of 64.37 kg CH<sub>4</sub> AU<sup>-1</sup> yr<sup>-1</sup>), because manure in deep-pits with long storage periods is conducive to generation of CH<sub>4</sub> due to anaerobic conditions. The pull-plug mode with manure regularly removed has the next highest CH<sub>4</sub> EF of 47.09 kg CH<sub>4</sub> AU<sup>-1</sup> year<sup>-1</sup>. In comparison, CH<sub>4</sub> emissions for separation mode are much lower with an EF of 10.93 kg CH<sub>4</sub> AU<sup>-1</sup> yr<sup>-1</sup>. The bedding mode has comparatively the lowest CH<sub>4</sub> EF (10.63 kg CH<sub>4</sub> AU<sup>-1</sup> yr<sup>-1</sup>) but the highest N<sub>2</sub>O EF (4.70 kg N<sub>2</sub>O AU<sup>-1</sup> yr<sup>-1</sup>) due to the nitrification and denitrification processes, which are facilitated by the co-existence of aerobic and anaerobic areas in the continuously accumulating manure on

the animal house floor. For NH<sub>3</sub> emissions, the bedding mode shows the lowest median value of 8.05 kg NH<sub>3</sub> AU<sup>-1</sup> yr<sup>-1</sup>; whereas for deep-pit, pull-plug and separation modes, the median NH<sub>3</sub> EFs are higher, in the range of 11.99-14.98 kg NH<sub>3</sub> AU<sup>-1</sup> yr<sup>-1</sup>. Therefore, changing manure cleaning system can achieve high CH<sub>4</sub> mitigation effect; with the CH<sub>4</sub> EF being reduced by 83% when the in-house manure collection method was changed from deep-pit method to solid-liquid separation method. Meanwhile, the difference of floor system (solid floor without urine drainage vs. slatted concrete floor) used in dairy house seemed have no significant influence on CH<sub>4</sub> emission (Pereira et al., 2011). However, the slatted floor can reduce NH<sub>3</sub> emission by 16-29% when compared with the solid floor. The reduced surface area of slatted floor caused the lower NH<sub>3</sub> emission. However, as the N<sub>2</sub>O emission was quite low being lower than 0.1% of the total N deposited on the floors, it's hard to compare the N<sub>2</sub>O emission from the two systems (Pereira et al., 2011).

#### 1.1.3 Litter additives

Many different kind of litter additives were proved to be quite efficient in reducing the NH<sub>3</sub> emissions in animal house, including alum (Eugene et al., 2015), Na bisulfate (PLT) (Shah et al., 2014), zeolite (Schneider et al., 2016). The mechanisms including reducing manure pH or adsorbing the polluted gases. However, Eugene et al. (2015) reported the concentrations and emissions of nitrous oxide (N<sub>2</sub>O) and methane (CH<sub>4</sub>) from the alum-treated house were not significantly different from the untreated house.

#### 1.1.4 biofilter used for exhausted air

Biofilters are widely used to reduce NH<sub>3</sub>, H<sub>2</sub>S, and to reduce CH<sub>4</sub> emissions of the exhausted air from the animal house (Gerber et al., 2013). The commonly used substrate materials for biofilters include manure compost, peat, wood chips, as well as some inorganic substrate materials added with microbial agents. Janni et al. (2014) found that the use of biofilter to treat the discharged gas outside the deep pit can reduce CH<sub>4</sub> emissions by 8-35%. Akdeniz et al. (2012) found that the CH<sub>4</sub> emission from the exhaust air of the pig house can be reduced by 22-28%. Meanwhile, methane removal up to 85% could be achieved in the experimental biofilter for treating air from the headspace of a covered 6 m<sup>3</sup> liquid manure storage (Melse and van der Werf, 2005). A large number of methane oxidizing bacteria are generally accumulated on biofilter, so CH<sub>4</sub> in the gas is oxidized and converted into CO<sub>2</sub> when passing through the biofilter

(Melse and Ogink, 2005). However, a large amount of  $\text{NH}_3$  also existed in the exhaust air. The biofilter can reduce  $\text{NH}_3$  emissions effectively through nitrification and denitrification, and the N element in  $\text{NH}_3$  would be partly converted into  $\text{N}_2\text{O}$  and  $\text{N}_2$ . Therefore, most studies have also found that the use of biofilters can increase  $\text{N}_2\text{O}$  emissions by 12-81% (Akdeniz et al., 2012; Janni et al., 2014). Melse et al. (2012) demonstrated that for a full-scale packed-bed biotrickling filter, 5% of the removed  $\text{NH}_3\text{-N}$  was converted into  $\text{N}_2\text{O-N}$  in the biofilter and emitted as  $\text{N}_2\text{O}$ .

## **1.2 Outdoor slurry storage**

For  $\text{CH}_4$  mitigation from slurry storage, cover, acidification and anaerobic digestion are the technologies which are widely used in practice or studied. Meanwhile, other possible measures include cooling and solid-liquid separation.

### **1.2.1 Covers**

Different types of covers can be used to reduce the gas emissions from animal slurry storage. However, the difference in the cover material, cover thickness, the climate condition and the using time may all influence the  $\text{CH}_4$  mitigation effect. It is generally believed that the use of plastic film has a constant effect on  $\text{CH}_4$  (Guarino et al., 2006), as plastic covering with secure sealing characteristics could help to avoid gas emissions. Straw is usually rich and cheap, and was widely studied to evaluate its effect on gas mitigation. Hansen et al. (2009) reported straw can reduce the  $\text{CH}_4$  emission by 26-50%, and Clemens et al. (2006) also reported the  $\text{CH}_4$  mitigation efficiency can reach 12-27% for dairy slurry storage with straw cover. However, some studies also reported the straw cover increased the  $\text{CH}_4$  emission by 3-76% (Berg et al., 2006; Guarino et al., 2006; Petersen et al., 2013; Amon et al., 2006; Clemens et al., 2006). It's reported that the decomposition of straw, if used for a prolonged period, may serve as an additional carbon source for methanogens, thus increasing  $\text{CH}_4$  emission (Amon et al., 2006). Different kind of granules can also serve as cover, and the granule covers can reduce  $\text{CH}_4$  emission by 9-14% (Berg et al., 2006; Guarino et al., 2006; Li, 2008). Oil can also be cover, and the floating oil can reduce  $\text{CH}_4$  emission by 10-13% (Guarino et al., 2006). Regardless of the varied materials, all the covers are efficient in  $\text{NH}_3$  mitigation, with the mitigation efficiency (ME) being usually higher than 50%. However, the use of straw cover and granule cover may usually increase  $\text{N}_2\text{O}$  emissions, due to nitrification

and denitrification processes occurring within the cover (Hansen et al., 2009; Berg et al., 2006). Although the cost of plastic film is higher than that of straw, its service life can be 20-25 times that of straw (Hörnig et al., 1999).

### 1.2.2 Slurry acidification

Acidification of animal slurry is a widely recognized  $\text{NH}_3$  reduction measure (Kai, 2008). Meanwhile, slurry acidification has also been shown to be effective in reducing  $\text{CH}_4$  emissions. Shin et al. (2019) adjusted the slurry pH at 5.0-7.0 using  $\text{H}_2\text{SO}_4$  solution, and found the  $\text{CH}_4$  emission can be reduced by 51-97%. Misselbrook et al. (2016) reported that compared with non-acidification conditions, acidifying dairy slurry to pH 5.5 can reduce  $\text{CH}_4$  emission by 63-90%. Wang et al. (2014a) reported that comparing with control, adjustment of initial pH of digested pig slurry to 5.5 significantly reduced  $\text{CH}_4$  emissions by 80.8% and  $\text{NH}_3$  emissions by 40.2%, but increased  $\text{H}_2\text{S}$  emissions by 11,324%; acidification with pH adjusted to 6.5 reduced  $\text{CH}_4$  emissions by 31.2%, but did not affect  $\text{NH}_3$  and  $\text{H}_2\text{S}$  emissions significantly. Besides the commonly used  $\text{H}_2\text{SO}_4$ , other types of acid were also tried to reduce the operation risk, including lactic acid, citric acid (Im et al., 2021),  $\text{FeCl}_3$ , and even some organic material which can cause self-acidification in the slurry, such as spent brewers' grain, sugarbeet (Kavanagh et al., 2021; Bastami et al., 2016). They are all proven to be effective in reducing  $\text{CH}_4$  emission, with the reported ME being in the range of 6-99% (Im et al., 2021; Kavanagh et al., 2021; Bastami et al., 2016). After the slurry is acidified to pH 5.5-6.0, the activity of internal methanogens is greatly reduced, and thus the methane mitigation effect is obtained (Shin et al., 2019). Ottosen et al. (2009) found that by adjusting the slurry pH to 5.5, the oxygen consumption rate and methanogenic activity in pig slurry decreased by 98%. However, the effect of acidification on  $\text{N}_2\text{O}$  is not clear. Berg et al. (2006) reported that the combination of acidifying slurry to pH 6.0 and perlite coverage can reduce  $\text{NH}_3$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  by 92%, 87% and 100%, respectively. In the same study, Berg et al. (2006) proved that the use of coverage alone would increase  $\text{N}_2\text{O}$  emission by 114%-152%, so it can be inferred that the addition of lactic acid is the main factor causing  $\text{N}_2\text{O}$  emission reduction.

### 1.2.3 Slurry Cooling

Since methanogens are extremely sensitive to temperature, cooling is an effective mean to reduce  $\text{CH}_4$  emissions (Umetsu et al., 2005; Petersen et al., 2013). Wang et al.

(2016) compared the emission characteristics of CH<sub>4</sub>, N<sub>2</sub>O and NH<sub>3</sub> during slurry storage at 5-25°C, and found that CH<sub>4</sub> emission is very small when the temperature is lower than 15°C (Wang et al., 2016). Lowering slurry storage temperature from 25°C to 20°C can reduce CH<sub>4</sub> emissions by 15%, while further reducing the temperature to lower than 15°C can reduce CH<sub>4</sub> emissions by 84-93% (Wang et al., 2016). Misselbrook et al. (2016) compared the CH<sub>4</sub> emission from pig and dairy slurry under cool, temperate and warm conditions, and found the CH<sub>4</sub> emission under cool condition was 98-99% lower than that under warm condition. In low temperature regions, frequent removal of manure from the house to the outside is a low-cost CH<sub>4</sub> mitigation option (Sommer et al., 2013; Petersen et al., 2013; Groenestein et al., 2012). As there are abundant methanogens existed in the fresh biogas slurry, the rapid cooling will reduce CH<sub>4</sub> emission as the methanogens usually produce CH<sub>4</sub> under mesophilic condition. Therefore, Sommer et al. (2000) suggested that the fresh biogas slurry should be cooled before it was stored in the storage tank to reduce CH<sub>4</sub> emissions. Temperature is one of the most important factors in determining NH<sub>3</sub> emission. Cooling slurry from 25 °C to 5 °C can reduce NH<sub>3</sub> emission by 60% (Dinuccio et al., 2008). N<sub>2</sub>O comes from nitrification and denitrification, and the activity of nitrifier and denitrifier was also affected by temperature. It's found that temperature higher than 25°C may stimulate N<sub>2</sub>O emission (Wang et al., 2016). Therefore, cooling can also reduce N<sub>2</sub>O emissions effectively.

#### 1.2.4 Anaerobic fermentation

Biogas recovery and utilization exhibited a high GHG mitigation potential. The substitution of the fossil fuel with the generated CH<sub>4</sub> is accounted into the GHG mitigation effect. For the CH<sub>4</sub> emission during the anaerobic fermentation process, approximately 10% of the CH<sub>4</sub> generated from biogas digesters may subsequently leak to the air according to 2006 IPCC guideline. Meanwhile, CH<sub>4</sub> loss from digestate storage is not negligible, and 5-15% additional biogas yield from digestate storage has been reported. All of these emissions should be taken into account when assessing the mitigation effect of biogas digesters. Unfortunately, there is no literature reporting a direct comparison of biogas digester vs. the baseline scenario. Therefore, the difference between the CH<sub>4</sub> emission from the biogas slurry storage and the raw slurry storage was regarded as the CH<sub>4</sub> mitigation effect in this study. As most of the organic matter has been decomposed during the anaerobic digestion process, there is always low

organic matter being reserved in the biogas slurry. Therefore, the CH<sub>4</sub> emission from the biogas slurry was 67-99% lower than the raw slurry (Wang et al., 2014b; Amon et al., 2006; Maldaner et al., 2018; VanderZaag et al., 2018), although there is also one study reporting that the CH<sub>4</sub> emission may be increased by 234% when compared with the raw slurry (Rodhe et al., 2015). The NH<sub>3</sub> emission in biogas slurry was 41-45% lower than raw slurry (Wang et al., 2014; Amon et al., 2006). However, the biogas slurry usually emitted much higher N<sub>2</sub>O than the raw slurry, being 76-362% (Wang et al., 2014b; Amon et al., 2006), mainly caused by the decreased COD/N in the digested slurry (Wang et al., 2014b).

### 1.2.5 Other mitigation options

It's reported that the CH<sub>4</sub> emission from the aerated slurry can be 57-87% lower than the control (Amon et al., 2006; Loyon et al., 2007). However the aeration procedure may increase N<sub>2</sub>O emission and NH<sub>3</sub> emission by 144% and 409%, respectively (Amon et al., 2006). Mechanical separation of slurry into its solid and liquid components is widely used to ease the transport of nutrients surplus outside livestock dense areas towards livestock-free plantations (Maffia et al., 2020). The solid-liquid separation caused the CH<sub>4</sub> emission from the liquid part being 42-81% lower than the raw slurry (Amon et al. 2006; VanderZaag et al., 2018). However, the NH<sub>3</sub> emission from the liquid part was much higher than the raw slurry (Amon et al. 2006).

## 1.3 Manure compost

For manure compost, the possible CH<sub>4</sub> mitigation measures include compost additive, compost biofilter. Meanwhile, cover was also tried to reduce the gas emissions from the compost process.

### 1.3.1 Compost additive

The compost additives were usually used for reducing NH<sub>3</sub> emission hence improving the retention of N in the final compost product. The commonly used additives include some inorganic additives, such as modified red mud, superphosphate, modified forsterite, and also include biochar, microbial additives, etc. Hao et al. (2005) reported that adding phosphogypsum in beef cattle manure composting with an amount of 10-30% of the dry weight can reduce CH<sub>4</sub> emissions from 15.36 kg C/Mg to 0.44-2.83 kg C/Mg, and higher mitigation was achieved with higher additive

amount. Phosphogypsum brings in  $\text{SO}_4^{2-}$  ions, and  $\text{SO}_4^{2-}$  ions are toxic to methanogens, leading to a reduction in methane production (Hao et al., 2005; Shin et al., 2019). Different from  $\text{CH}_4$  emission, adding phosphogypsum to cattle manure compost may increase  $\text{N}_2\text{O}$  emissions by 78-156% (Hao et al., 2005). As phosphogypsum addition can significantly reduce  $\text{NH}_3$  emissions, it may lead to more N being retained in the compost pile, providing more N element for  $\text{N}_2\text{O}$  generation from nitrification and denitrification. Yang et al. (2015) reported that the superphosphate addition in compost can reduce  $\text{CH}_4$  emission by 32-84%. However, Luo et al. (2012) found that adding superphosphate to pig manure compost may increase  $\text{CH}_4$  by 4-15%. Biochar addition can reduce  $\text{CH}_4$  emission from poultry manure compost process by 78-84% (Chowdhury et al., 2014; Agyarko-Mintah et al., 2017). Some studies reported a lower  $\text{CH}_4$  mitigation effect with biochar addition, being 6-26% (Chen et al., 2017; Chen 2016). Biochar was usually deemed with relatively stable properties and it is difficult to be degraded by microorganisms as a C source; in addition, the addition of biochar may increase the pores of the compost pile, leading to a better aerobic condition, thus contributing to reducing  $\text{CH}_4$  emissions (Agyarko-Mintah et al., 2017). However, if the biochar is added as powder, it may cause the increase of  $\text{CH}_4$  emission by 57% because of the decreased aerobic condition (He et al., 2018). The use of bulking agent such as straw can also help reduce  $\text{CH}_4$  emission, being in the range of 45-74% as these materials help reduce the density but improve the porousness of the pile; meanwhile, the  $\text{N}_2\text{O}$  can be reduced by 11-64% (Yamulki et al., 2006; Maeda et al., 2013).

### 1.3.2 Compost biofilter

Biofilter can also be used to treat the exhaust air from composting process. Zhou (2017) found the biofilter for treating the exhaust air from pig manure compost can reduce  $\text{CH}_4$  emission by 85%. It can also reduce  $\text{NH}_3$  emission efficiently, with the ME being in the range of 81-100% (Park 2002; Xue et al., 2010).

### 1.3.3 Compost cover

With the plastic foil cover, the  $\text{CH}_4$  emission from compost was increased by 32-57% because of the worse aerobic condition (Jiang et al., 2011, 2013). However, cover can reduce the  $\text{N}_2\text{O}$  and  $\text{NH}_3$  emission, being in the range of 10-39% and 3-35%, respectively (Jiang et al., 2011, 2013).

## **1.4 Manure stockpile**

The effect of cover on CH<sub>4</sub> emission from manure stockpile was not consistent. Hansen et al. (2006) reported that the Polyethylene cover can reduce CH<sub>4</sub> emission from pig manure stockpile by 88%. Naylor et al. (2016) reported the impermeable cover caused CH<sub>4</sub> emission increasing by 6300% from poultry manure stockpile. For straw cover, it may increase CH<sub>4</sub> emission by 6-44% (Zhu et al., 2014, 2017). Cover showed a consistent mitigation effect on NH<sub>3</sub> emission. The impermeable cover can help reduce N<sub>2</sub>O emission. However, the straw cover may sometimes increase the N<sub>2</sub>O emission (Zhu et al., 2014). The carbon input sourcing from the straw and sawdust stimulated the nitrification and denitrification process in the pile, and may trigger the increased N<sub>2</sub>O emissions.

## **1.5 Land application**

For CH<sub>4</sub> mitigation from rice paddy field, it's widely found that replacing chemical fertilization with manure application usually increases the CH<sub>4</sub> emission (Pandey et al., 2014; Yang et al., 2015). The CH<sub>4</sub> emission from the cropland was usually low as the cropland was usually seen as carbon sink. The mitigation options used in cropland usually aimed at lowering NH<sub>3</sub> and N<sub>2</sub>O emission, thus lowering the odor emission, GHG emission and enhancing the N retention in the field. Changing slurry broadcast to slurry injection can reduce NH<sub>3</sub> emission by 78%, and the CH<sub>4</sub> emission was also reduced by 39% (Fangueiro et al., 2015). The NH<sub>3</sub> emission can be reduced by 97% if the applied slurry was being pre-acidified, however, the CH<sub>4</sub> emission may be increased by 18% (Fangueiro et al., 2015). Nitrification inhibitor (NI) additive was efficient in reducing the N<sub>2</sub>O emission, with the ME being 47-88% (Cahalan 2015; Tao et al., 2008; Minet et al., 2015). Meanwhile, Minet et al. (2015) reported the CH<sub>4</sub> emission may also be reduced by 42% by using NI additive. The effect of substituting the raw slurry with biogas slurry on gas emissions was not consistent (Clemens et al., 2006; Amon et al., 2006).

## **1.6 Selection of the most recommended technology for each stage**

Based on the literature review, the recommended CH<sub>4</sub> mitigation options in each stage were listed in Table 5. For the in-house stage, increasing the manure removal frequency is quite recommended in reducing CH<sub>4</sub> emission, with the mitigation

efficiency being in the range of 27-39%. However, its effect on N<sub>2</sub>O and NH<sub>3</sub> emission was not sure. Biofilter is useful in reducing the CH<sub>4</sub> emission from the exhaust air, with the CH<sub>4</sub> ME being 8-85%. Meanwhile, the NH<sub>3</sub> can also be mitigated simultaneously with the ME being 53-90%. However, the N<sub>2</sub>O may be increased because of the nitrification and denitrification mechanism in reducing NH<sub>3</sub> emission.

Acidification is quite recommended for reducing CH<sub>4</sub> emission from slurry storage, with the CH<sub>4</sub> ME being 31-99%, and the NH<sub>3</sub> can also be reduced by 49-84% by using this technology. However, its effect on N<sub>2</sub>O emission was not sure.

For manure compost, the biofilter can also be used to treat the exhaust air of the compost pile, with CH<sub>4</sub> emission being reduced by 85%. The NH<sub>3</sub> emission can be reduced by 81-100%, and the N<sub>2</sub>O emission may be increased. Meanwhile, biochar was supposed to be an effective additive for reducing the CH<sub>4</sub> emission from compost process, with the ME being 6-88%. The NH<sub>3</sub> can also be reduced by 9-43% by adding biochar. In addition, biochar addition may also decrease the N<sub>2</sub>O emission by 11-75%.

For manure stockpile, no mitigation option was proven to be usually efficient in CH<sub>4</sub> mitigation. But the used cover can usually reduce NH<sub>3</sub> emission.

For land-application, crop-land was usually seen as CH<sub>4</sub> sink while rice paddy field may be an important source of CH<sub>4</sub> emission. However, there is quite little literature being available for studying the CH<sub>4</sub> mitigation options. Therefore, no mitigation option for CH<sub>4</sub> mitigation was recommended in this stage.

Table 1 Effects of different management options on CH<sub>4</sub> emissions together with N<sub>2</sub>O and NH<sub>3</sub> emission from pig manure management

Stage	Mitigation option		CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
Housing	diet additive	with CaSO <sub>4</sub> -H <sub>3</sub> PO <sub>4</sub>	(-14%)			Kim et al., 2004
		with CaCl <sub>2</sub>	73%		(-37%)	Eriksen et al., 2014
	biofilter		(-80%)~(-85%)			Melse and van der Werf, 2005
			(-8%)~(-35%)	12% ~ 81%	(-53%)~(-86%)	Janni et al., 2014
			(-22%)~(-28%)	12%~29%	(-61%)~(-86%)	Akdeniz et al., 2012
			(-31%)	(-16%)	(-90%)	Hood et al., 2015
			(-18%)			Girard et al., 2011
	Increase manure cleaning frequency		(-39%)			Ulens et al., 2014
			(-27%)		25%	Wang et al., 2017
	Slurry storage	straw cover		49%	468%	(-75%)
			7~60%		(-14%)~(-84%)	Guarino et al., 2006
			(-8%)~(-28%)		(-1%)~(-86%)	Guarino et al., 2006
			(-26%)~(-50%)	1438~10692%		Hansen et al., 2009
			(-9%)~(-29%)	39600%	(-50%)~(-89%)	Petersen et al., 2013
			76%		(-64%)	Petersen et al., 2013
oil cover			(-10%)~(-13%)		(-80%)~(-100%)	Guarino et al., 2006
					(-93%)	Portejoie et al., 2003
					(-50%)~(-85%)	Hornig et al., 1999
granule cover		Perlite cover	(-14%)	280%	(-90%)	Berg et al., 2006
		leca cover	0%	262%	(-83%)	Berg et al., 2006
		Expanded clay cover	(-9%)~(-17%)		(-17%)~(-75%)	Guarino et al., 2006
		Clay cover	(-14%)			Li, 2008
cooling			(-23%)~(-46%)			Groenestein et al., 2012
			(-61%)	no effect	(-60%)	Dinuuccio et al., 2008

Stage	Mitigation option		CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
			(-15%)~(-93%)	(-81%)~(-91%)	(-4%)~(-84%)	Wang et al., 2016
	acidification	H <sub>2</sub> SO <sub>4</sub>	(-31%)~(-81%)		(-40%)	Wang et al., 2014a
		H <sub>2</sub> SO <sub>4</sub>	(-44%)~(-94%)		(-49%)~(-84%)	Petersen et al., 2014
		H <sub>2</sub> SO <sub>4</sub>	(-51%)~(-97%)			Shin et al., 2019
		citric acid	(-85%)~(-99%)			Im et al., 2021
	aeration		(-87%)		(-96%)	Loyon et al., 2007
	anaerobic digestion		-99%	362%	(-45%)	Wang et al., 2014b
compost	additive	superphosphate	(-22%)~(-63%)	(-21%)~(-26%)	(-23%)~(-42%)	Luo et al., 2012
		superphosphate	4%~15%	(-14%)~(-32%)	(-56%)~(-64%)	Luo et al., 2012
		biochar (powder)	57%			He et al., 2018
		biochar (granular)	(-22%)			He et al., 2018
	cover	Plastic foil	32%~46%	(-10%)~(-39%)	(-3%)~(-35%)	Jiang et al., 2013
		Plastic foil	44%~57%	(-25%)~(19%)	(-10%)~(-20%)	Jiang et al., 2011
		biofilter		(-85%)		(-81%)~(-99%)
solid stockpile	cover	Polyethylene cover	(-88%)	(-99%)	(-12%)	Hansen et al., 2006

Table 2 Effects of different management options on CH<sub>4</sub> emissions together with N<sub>2</sub>O and NH<sub>3</sub> emission from dairy manure management

Stage	Mitigation option		CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
Housing	slatted vs solid		0%	(-9%)	(-16%)	Pereira et al., 2011
			4%	(-49%)	(-24%)	Pereira et al., 2011
			(-10%)	(-38%)	(-29%)	Pereira et al., 2011
slurry storage	cooling	to lower than 15 oC	(-43%)~(-100%)			Clemens et al., 2006
		to lower than 15 oC	(-98%)~(-99%)		(-48%)~(-68%)	Misselbrook et al., 2016
	cover	wood cover	(-14%)~(-16%)	(20%)~(-13%)	(-28%)~(-46%)	Clemens et al., 2006
		straw cover	3%	0-5%	(-20%)~(-44%)	Clemens et al., 2006

Stage	Mitigation option	CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
	straw cover	21%	109%	24%	Amon et al., 2006
	straw+wood cover	(-12%)~(-27%)	(1%)~(-15%)	(-21%)~(-65%)	Clemens et al., 2006
	calcium carbonate + hydrophobic silica cover	102%-117%			Sakamoto et al., 2008
	superphosphate + hydrophobic silica cover	(-65%)~(-45%)			Sakamoto et al., 2008
	Biofiltration	(-90%)~(-99%)		(-69%)	Pratt et al., 2012
Acidification	H <sub>2</sub> SO <sub>4</sub>	(-63%)~(-90%)		(-56%)~(-99%)	Misselbrook et al., 2016
	H <sub>2</sub> SO <sub>4</sub>	1.0%	4%	(-69%)	Owusu-Twum et al. 2017
	H <sub>2</sub> SO <sub>4</sub>	(-69%)~(-84%)			Habtewold et al., 2018
	H <sub>2</sub> SO <sub>4</sub>	(-68%)		(-62%)	Sommer et al., 2017
	H <sub>2</sub> SO <sub>4</sub>	(-67%)~(-87%)			Petersen et al., 2012
	H <sub>2</sub> SO <sub>4</sub>	(-87%)~(-89%)		(-41%)~(-53%)	Sokolov et al., 2020
	Brewing sugar	(-87%)~(-99%)			Bastami et al., 2016
	organic material	(-15%)~(-70%)		(-38%)~(-67%)	Kavanagh et al., 2021
	FeCl <sub>3</sub>	(-6%)~(-65%)		(-20%)~(-68%)	Kavanagh et al., 2021
	Aerated	(-57%)	144%	409%	Amon et al., 2006
	Solid-liquid separated	(-42%)	10%	698%	Amon et al., 2006
		(-81%)			VanderZaag et al., 2018
Anaerobic digestion	biogas slurry vs raw slurry	(-67%)	76%	(-41%)	Amon et al., 2006
	biogas slurry vs raw slurry	234%	occurred		Rodhe et al., 2015
	biogas slurry vs raw slurry	(-85%)			Maldaner et al., 2018
	biogas slurry vs raw slurry	(-59%)			VanderZaag et al., 2018
Compost	Mixed with old slurry	129%	(-11%)	(-69%)	Wood et al., 2014
	additive straw	(-50%)~(-45%)	(-11%)~(-42%)		Yamulki et al., 2006
	additive Bulking agent	(-74%)	(-64%)	127%	Maeda et al., 2013
Land Application	Acidification	18%	(-3%)	(-97%)	Fangueiro et al., 2015

Stage	Mitigation option		CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
					(-96%)	Owusu-Twum et al.2017
	NI additive		(-42%)	(-88%)		Minet et al., 2015
	pre-digested	digested vs raw	(-41%)	5%	12%	Clemens et al., 2006
		digested vs raw	54%	(-29%)	18%	Amon et al., 2006
	pre-seperated	seperated vs raw	1415%	68%	(-59%)	Amon et al., 2006
					(-49%)	Owusu-Twum et al.2017
	Injection		(-39%)	(-13%)	(-78%)	Fangueiro et al., 2015

Table 3 Effects of different management options on CH<sub>4</sub> emissions together with N<sub>2</sub>O and NH<sub>3</sub> emission from beef cattle manure management

Stage	Mitigation option		CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference
Feed	low CP		no significant influence			Menezes et al., 2016
Slurry	additive	effective microorganisms	(-15%)~(27%)			Bastami et al., 2016
		sugar	(-89%)~(-99%)			Bastami et al., 2016
		sugar + effective microorganisms	(-85%)~(-99%)			Bastami et al., 2016
Compost	additive	phosphogypsum	(-82%)~(-97%)	78%~157%		Hao et al., 2005
		biofilter			(-81%)~(-100%)	Park et al., 2002
		biofilter			(-95%)	Xue et al., 2010
Stockpile	cover	straw cover	5.7~21%	24.9~30.7%		Zhu et al., 2014
		sawdust cover	17.6~43.8%	(-20%)~(25%)	(-15%)~(-58%)	Zhu et al., 2017
	cooling		(-94%)			Mathot et al., 2012
	straw addition		(-45%)~(-51%)			Yamulki et al., 2006
Land application	additive	nitrification inhibitor	no significant influence	(-47%)~(-71%)		Cahalan et al., 2015
		nitrification inhibitor		(-73%)~(-79%)	5%-100%	Tao et al., 2008

Table 4 Effects of different management options on CH<sub>4</sub> emissions together with N<sub>2</sub>O and NH<sub>3</sub> emission from layer manure management

Stage	Mitigation option	CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>	Reference	
In house	LCP diet	(-26%)	(-7%)	(-27%)	Wiedemann et al., 2016	
	diet additive	emission reduction diet	(-10%)		(-27%)	Wu-Haan et al., 2007
		emission reduction diet	(-36%)		(-45%)	
		emission reduction diet	(-17%)		(-46%)	
	Litter additive	alum	7%		(-52%)	Eugene et al., 2015
		alum	28%		(-51%)	
alum		17%		(-40%)		
Compost	additive	biochar	(-11%)~(-26%)		(-9%)~(-25%)	Chen et al., 2017
		biochar	(-78%)~(-84%)	(-11%)~(-16%)	(-35%)~(-43%)	Chowdhury et al., 2014
		biochar	(-78%)~(-83%)	(-68%)~(-75%)		Agyarko-Mintah et al., 2017
		biochar	(-6%)~(-22%)		(-21%)~(-33%)	Chen 2016
		superphosphate	(-32%)~(-84%)	(-2%)~(-36%)	(-14%)~(-27%)	Yang et al., (2015)
Stockpile	cover	impermeable cover	6300%	(-100%)	(-88%)	Naylor et al., 2016

**Table 5 Best options for reducing CH<sub>4</sub> emission from each stage of animal manure management**

Stage	Mitigation option	CH <sub>4</sub>	N <sub>2</sub> O	NH <sub>3</sub>
In-house	frequent removal	(-27%)~(-39%)	not sure	not sure
Exhaust air	biofilter	(-8%)~(-85%)	increase	(-53%)~(-90%)
Slurry	Acidification	(-31%)~(-99%)	not sure	(-49%)~(-84%)
compost	biofilter	(-85%)	increase	(-81%)~(-100%)
	biochar additive	(-6%)~(-88%)	(-11%)~(-75%)	(-9%)~(-43%)
stockpile	cover	not sure	not sure	(-15%)~(-88%)

## 2 Integration of the technologies

Integrated technology can be used to achieve a better gas mitigation effect. For slurry storage, cooling and acidification can be combined to reach a better CH<sub>4</sub> mitigation effect. Im et al. (2020) reported that to achieve 70% CH<sub>4</sub> reduction, 1.6 kg H<sub>2</sub>SO<sub>4</sub>/ton pig slurry was needed in pig slurry acidification, which was decreased to 0.5 kg H<sub>2</sub>SO<sub>4</sub>/ton pig slurry by decreasing temperature from 35 °C to 25 °C. The CH<sub>4</sub> reduction efficiency of storage-temperature decrease was more effective in the acidified PS with low strength. Meanwhile, acidification can also be combined with the coverage. Berg et al. (2006) reported that the combination of acidifying slurry to pH 6.0 and perlite coverage can reduce CH<sub>4</sub>, N<sub>2</sub>O and NH<sub>3</sub> by 87%, 100% and 92%, respectively. Wang et al. (2021) also reported that for the combination of acidification and coverage technology by using H<sub>2</sub>SO<sub>4</sub>-modified expanded vermiculite on slurry surface, the CH<sub>4</sub> and NH<sub>3</sub> emissions can be reduced by 52% and 87%, respectively. The MEs of the combined technology were comparable with those achieved by acidifying slurry to pH 6.0, with the CH<sub>4</sub> and NH<sub>3</sub> being reduced by 50% and 90%.

For land-application, Fangueiro et al. (2018) reported that compared with slurry injection, band application of acidified slurry without soil incorporation reduced the N<sub>2</sub>O and CH<sub>4</sub> emissions by 65% and 40%, respectively. The lower CH<sub>4</sub> production after amendment with acidified slurry could be related to the lower soil pH that might have inhibited methanogenesis (Fangueiro et al., 2015).

For the whole MMC, Sajeev et al. (2017) reported that a shift from single-stage emission abatement options towards a whole-chain perspective is vital in reducing overall emissions. It was recommended that the best integrated CH<sub>4</sub> mitigation package used in animal MMC included frequent manure removal during housing stage, using acidification for manure treatment, and using covers during manure storage stage. No option was recommended for the land-application stage because of the almost nil CH<sub>4</sub> emission during this stage.

After a pre-screening of the best mitigation option in each stage by using meta-analysis, Wang et al. (2017) recommended the combined mitigation strategies used for reducing CH<sub>4</sub> emission from the whole pig MMC, including the in-house stage, manure storage and treatment stage. As usually no CH<sub>4</sub> emission occurred in the land-application stage, the mitigation package was targeted at the first two stages.

For the two liquid manure management chain, including deep-pit pig MMC and pull-plug pig MMC, the combination of low-CP diet, biofilter for treating the house exhaust air, and the slurry acidification can reduce 59-63% of the CH<sub>4</sub> emission from the whole MMC. Although N<sub>2</sub>O emission was increased by using biofilter, the total GHG was still reduced by 47-51% as CH<sub>4</sub> was the major GHG contributor in liquid manure management system. Meanwhile, the NH<sub>3</sub> emission of the chain can also be reduced by 38%-40% (Wang et al., 2017).

For bedding system, the combined use of LCP diet, biofilter for treating the house exhaust air and compost additive can reduce the whole chain CH<sub>4</sub> emission by 23%. However, the total GHG can only be reduced by 10%, as the use of biofilter may increase N<sub>2</sub>O emission. Meanwhile, the whole chain NH<sub>3</sub> can be reduced by 41% with this mitigation package (Wang et al., 2017).

For solid-liquid separation system, the manure was separated into solid part and liquid part. The combination of low CP diet, biofilter for treating the house exhaust air, and additive used for the solid compost part and liquid slurry acidification can achieve a whole chain CH<sub>4</sub> mitigation efficiency of 28.6%. The total GHG of the whole MMC can be reduced by 20%. The whole chain NH<sub>3</sub> can be reduced by 43% (Wang et al., 2017).

## References

- 1) Agyarko-Mintah, E., Cowie, A., Van, Z.L., Singh, B.P., Smillie, R., Harden, S., Fornasier, F., 2017. Biochar lowers ammonia emission and improves nitrogen retention in poultry litter composting. *Waste Manag.* 61, 129-137.
- 2) Akdeniz, N., Janni, K.A., 2012 Full-scale biofilter reduction efficiencies assessed using portable 24-hour sampling units. *J. Air Waste Manage. Assoc.* 62(2), 170-182.
- 3) Amon, B., Kryvoruchko, V., Amon, T., Zechmeister-Boltenstern, S., 2006. Methane, nitrous oxide and ammonia emissions during storage and after application of dairy cattle slurry and influence of slurry treatment. *Agriculture, ecosystems & environment*, 112(2-3), 153-162.
- 4) Bastami, M.S. B., Jones, D. L., Chadwick, D. R., 2016. Reduction of methane emission during slurry storage by the addition of effective microorganisms and excessive carbon source from brewing sugar. *J. Environ. Qual.* 45(6), 2016-2022
- 5) Berg, W., Brunsch, R., Pазsiczki, I., 2006. Greenhouse gas emissions from covered slurry compared with uncovered during storage. *Agric. Ecosyst. Environ.* 112(2), 129-134.
- 6) Cahalan, E., Ernfors, M., Müller, C., Devaney, D., Laughlin, R. J., Watson, C. J., Hennessy, D., Grant, J., Khalil, M. I., McGeough, K. L., Richards, K. G., 2015. The effect of the nitrification inhibitor dicyandiamide

- (DCD) on nitrous oxide and methane emissions after cattle slurry application to Irish grassland. *Agr. Ecosyst. Environ.* 199, 339-349.
- 7) Chen W. 2016. A harmful gases mitigation technology study in layer manure composting based on biochar selection and modification. HuaNan Agriculture University. Guangzhou. (in Chinese with English abstract)
  - 8) Chen, W., Liao, X., Wu, Y., Liang, J.B., Mi, J., Huang, J., Zhang, H., Wu, Y., Qiao, Z., Li, X., Wang, Y., 2017. Effects of different types of biochar on methane and ammonia mitigation during layer manure composting. *Waste Manage.* 61, 506-515.
  - 9) Chowdhury, M.A., de Neergaard, A., Jensen, L.S., 2014. Potential of aeration flow rate and bio-char addition to reduce greenhouse gas and ammonia emissions during manure composting. *Chemosphere* 97, 16-25.
  - 10) Clemens J, Trimborn M, Weiland P, Weiland, P., Amon, B., 2006. Mitigation of greenhouse gas emissions by anaerobic digestion of cattle slurry. *Agriculture Ecosystems & Environment*, 112(2–3):171-177.
  - 11) Dinuccio, E., Berg, W., Balsari, P., 2008. Gaseous emissions from the storage of untreated slurries and the fractions obtained after mechanical separation. *Atmospheric Environment*, 42(10): 2448-2459.
  - 12) Eriksen, J., Nørgaard, J. V., Poulsen, H. D., Poulsen, H. V., Jensen, B. B., & Petersen, S. O. (2014). Effects of acidifying pig diets on emissions of ammonia, methane, and sulfur from slurry during storage. *Journal of environmental quality*, 43(6), 2086-2095.
  - 13) Eugene, B., Moore, P.A., Li, H., Miles, D., Trabue, S., Burns, R., Buser, M., 2015. Effect of alum additions to poultry litter on in-house ammonia and greenhouse gas concentrations and emissions. *J. Environ. Qual.* 44, 1530-1540.
  - 14) Fangueiro, D., Pereira, J. L., Fraga, I., Surgy, S., Vasconcelos, E., Coutinho, J., 2018. Band application of acidified slurry as an alternative to slurry injection in a Mediterranean double cropping system: agronomic effect and gaseous emissions. *Agriculture, Ecosystems & Environment*, 267, 87-99.
  - 15) Fangueiro, D., Surgy, S., Fraga, I., Cabral, F., Coutinho, J., 2015. Band application of treated cattle slurry as an alternative to slurry injection: implications for gaseous emissions, soil quality and plant growth. *Agric. Ecosyst. Environ.* 211, 102–111.
  - 16) Gerber, P.J., Hristov, A.N., Henderson, B., Makkar, H., Oh, J., Lee, C., Meinen, R., Montes, F., Ott, T., Firkins, J., Rotz, A., Dell, C., Adesogan, A.T., Yang, W.Z., Tricarico, J.M., Kebreab, E., Waghorn, G., Dijkstra, J., Oosting, S., 2013. Technical options for the mitigation of direct methane and nitrous oxide emissions from livestock: a review. *Animal*, 7(s2):220-234.
  - 17) Girard, M., Ramirez, A. A., Buelna, G., Heitz, M., 2011. Biofiltration of methane at low concentrations representative of the piggery industry—influence of the methane and nitrogen concentrations. *Chemical Engineering Journal*, 168(1), 151-158.
  - 18) Groenestein, K., Mosquera, J., Van der Sluis, S., 2012. Emission factors for methane and nitrous oxide from manure management and mitigation options. *J. Integr. Environ. Sci.* 9(sup1), 139-146.
  - 19) Guarino, M., Fabbri, C., Brambilla, M., Valli, L., Navarotto, P., 2006. Evaluation of simplified covering systems to reduce gaseous emissions from livestock manure storage. *T. ASAE.* 49(3), 737-747.
  - 20) Habtewold, J., Gordon, R., Sokolov, V., VanderZaag, A., Wagner-Riddle, C., Dunfield, K., 2018. Reduction in methane emissions from acidified dairy slurry is related to inhibition of Methanosarcina species. *Frontiers in microbiology*, 9, 2806.
  - 21) Hansen, M.N., Henriksen, K., Sommer, S.G., 2006. Observations of production and emission of greenhouse gases and ammonia during storage of solids separated from pig slurry: Effects of covering. *Atmos. Environ.* 40(22), 4172-4181.
  - 22) Hansen, R.R., Nielsen, D.A., Schramm, A., Nielsen, L.P., Revsbech, N.P., Hansen, M.N., 2009. Greenhouse gas microbiology in wet and dry straw crust covering pig slurry. *J. Environ Qual.* 38(3), 1311-1319.

- 23) Hao, X., Larney, F. J., Chang, C., Travis, G. R., Nichol, C. K., Bremer, E., 2005. The effect of phosphogypsum on greenhouse gas emissions during cattle manure composting. *J. Environ. Qual.* 34(3),774-781.
- 24) He, X., Yin, H., Sun, X., Han, L., Huang, G., 2018. Effect of different particle-size biochar on methane emissions during pig manure/wheat straw aerobic composting: Insights into pore characterization and microbial mechanisms. *Bioresource technology*, 268, 633-637.
- 25) Hood, M. C., Shah, S. B., Kolar, P., Li, L. W., Stikeleather, L., 2015. Biofiltration of ammonia and GHGs from swine gestation barn pit exhaust. *Transactions of the ASABE*, 58(3), 771-782.
- 26) Hörnig, G., Türk, M., Wanka, U., 1999. Slurry covers to reduce ammonia emission and odour nuisance. *J. Agr. Eng. Res.* 73(2),151-157.
- 27) Im, S., Mostafa, A., Kim, D. H., 2021. Use of citric acid for reducing CH<sub>4</sub> and H<sub>2</sub>S emissions during storage of pig slurry and increasing biogas production: Lab-and pilot-scale test, and assessment. *Science of the Total Environment*, 753, 142080.
- 28) Im, S., Mostafa, A., Shin, S. R., Kim, D. H., 2020. Combination of H<sub>2</sub>SO<sub>4</sub>-acidification and temperature-decrease for eco-friendly storage of pig slurry. *Journal of Hazardous Materials*, 399, 123063.
- 29) Janni, K., Jacobson, L., Hetchler, B., Oliver, J., Johnston, L., 2014. Semi-continuous air sampling versus 24-hour bag samples to evaluate biofilters on a swine nursery in warm weather. *T. ASABE.* 57(5), 1501-1515.
- 30) Jiang, T., Schuchardt, F., Li, G., 2011. Effect of turning and covering on greenhouse gas and ammonia emissions during the winter composting. *T. CSAE.* 27(10): 212-217. (in Chinese with English abstract)
- 31) Jiang, T., Schuchardt, F., Li, G.X., Guo, R., Luo, Y.M., 2013. Gaseous emission during the composting of pig feces from Chinese Ganqinfen system. *Chemosphere*, 90(4),1545-1551.
- 32) Kai, P., Pedersen, P., Jensen, J., Hansen, M.N., Sommer, S.G., 2008. A whole-farm assessment of the efficacy of slurry acidification in reducing ammonia emissions. *Eur. J. Agron.* 28(2), 148-154.
- 33) Kavanagh, I., Fenton, O., Healy, M. G., Burchill, W., Lanigan, G. J., Krol, D. J., 2021. Mitigating ammonia and greenhouse gas emissions from stored cattle slurry using agricultural waste, commercially available products and a chemical acidifier. *Journal of Cleaner Production*, 294, 126251.
- 34) Kim, I.B. Kim, P.R. Ferket, W.J. Powers, H.H. Stein, T.A.T. van Kempen, 2004. Effects of different dietary acidifier sources of calcium and phosphorus on ammonia, methane, and odorant emission from growing-finishing pigs Asian-Aust. *J. Anim. Sci.*, 17, 1131-1138
- 35) Li, N. 2008. GHG emission from slurry storage of swine farm. Master thesis. Chinese Academy of Agricultural sciences, Beijing, China (in Chinese with english abstract).
- 36) Loyo, L., Guiziou, F., Béline, F., Peu, P., 2007. Gaseous emissions (NH<sub>3</sub>, N<sub>2</sub>O, CH<sub>4</sub> and CO<sub>2</sub>) from the aerobic treatment of piggery slurry—comparison with a conventional storage system. *Biosystems engineering*, 97(4), 472-480.
- 37) Luo, Y., Li, G.X., Wang, K., Jiang, T., Luo, W.H., 2012. Effects of additive superphosphate on NH<sub>3</sub>, N<sub>2</sub>O and CH<sub>4</sub> emissions during pig manure composting. *T. CSAE.* 28(22), 235-242.(in Chinese with English abstract)
- 38) Maeda, K., Hanajima, D., Morioka, R., Toyoda, S., Yoshida, N., Osada, T., 2013. Mitigation of greenhouse gas emission from the cattle manure composting process by use of a bulking agent. *Soil Science and Plant Nutrition*, 59(1), 96-106.
- 39) Maffia, J., Gioelli, F., Rollè, L., Airoldi, G., Balsari, P., & Dinuccio, E., 2020. Addition of powdery sulfur to pig slurry to reduce NH<sub>3</sub> and GHG emissions after mechanical separation. In *2020 IEEE International Workshop on Metrology for Agriculture and Forestry (MetroAgriFor)* (pp. 49-52). IEEE.
- 40) Maldaner, L., Wagner-Riddle, C., VanderZaag, A. C., Gordon, R., Duke, C., 2018. Methane emissions from storage of digestate at a dairy manure biogas facility. *Agricultural and forest meteorology*, 258, 96-107.

- 41) Mathot, M., Decruyenaere, V., Stilmant, D., Lambert, R., 2012. Effect of cattle diet and manure storage conditions on carbon dioxide, methane and nitrous oxide emissions from tie-stall barns and stored solid manure. *Agriculture, Ecosystems & Environment*, 148, 134-144.
- 42) Melse, R., Ogink, N., 2005. Air scrubbing techniques for ammonia and odor reduction at livestock operations: Review of on-farm research in the Netherlands. *Trans. ASAE* 48, 2303-2313.
- 43) Melse, R.W., Ploegaert, J.P., Ogink, N.W., 2012. Biotrickling filter for the treatment of exhaust air from a pig rearing building: Ammonia removal performance and its fluctuations. *Biosys. Eng. 113(3):242-252*.
- 44) Melse, R.W., van der Werf, A.W., 2005. Biofiltration for mitigation of methane emission from animal husbandry. *Environ. Sci. Technol.* 39(14), 5460-5468.
- 45) Menezes, A. C. B., Valadares Filho, S. C., e Silva, L. F. C., Pacheco, M. V. C., Pareira, J. M. V., Rotta, P. P., Zanetti, D., Detmann, E., Silva, F. A. S., Godoi, L. A., Renno, L. N., 2016. Does a reduction in dietary crude protein content affect performance, nutrient requirements, nitrogen losses, and methane emissions in finishing Nellore bulls? *Agr. Ecosyst. Environ.* 223, 239-249.
- 46) Minet, E. P., Jahangir, M.M.R., Krol, D.J., Rochford, N., Fenton, O., Rooney, D., Richards, K. G., 2015. Amendment of cattle slurry with the nitrification inhibitor dicyandiamide during storage: A new effective and practical N<sub>2</sub>O mitigation measure for land spreading. *Agriculture Ecosystems & Environment*, 215, 68-75.
- 47) Misselbrook, T., Hunt, J., Perazzolo, F., Provolo, G., 2016. Greenhouse Gas and Ammonia Emissions from Slurry Storage: Impacts of Temperature and Potential Mitigation through Covering (Pig Slurry) or Acidification (Cattle Slurry). *Journal of Environmental Quality*, 45(5), 1520-1530.
- 48) Naylor, T.A., Wiedemann, S.G., Phillips, F.A., Warren, B., Mcgahan, E.J., Murphy, C.M., 2016. Emissions of nitrous oxide, ammonia and methane from Australian layer-hen manure storage with a mitigation strategy applied. *Anim. Production Sci.* 56, 1367-1375.
- 49) Ottosen, L. D., Poulsen, H. V., Nielsen, D. A., Finster, K., Nielsen, L. P., Revsbech, N. P., 2009. Observations on microbial activity in acidified pig slurry. *Biosystems Engineering*, 102(3), 291-297.
- 50) Owusu-Twum, M. Y., Polastre, A., Subedi, R., Santos, A. S., Ferreira, L. M. M., Coutinho, J., Trindade, H. , 2017. Gaseous emissions and modification of slurry composition during storage and after field application: Effect of slurry additives and mechanical separation. *Journal of environmental management*, 200, 416-422.
- 51) Pandey, A., Mai, V.T., Vu, D.Q., Bui, T.P.L., Mai, T.L.A., Jesen, L.S., de Neergaard, A., 2014. Organic matter and water management strategies to reduce methane and nitrous oxide emissions from rice paddies in Vietnam. *Agric. Ecosyst. Environ.* 196, 137-146.
- 52) Park, K. J., Choi, M. H., Hong, J. H., 2002. Control of composting odor using biofiltration. *Compost Sci. Util.* 10(4), 356-362.
- 53) Pereira, J., Fangueiro, D., Misselbrook, T. H., Chadwick, D. R., Coutinho, J., Trindade, H., 2011. Ammonia and greenhouse gas emissions from slatted and solid floors in dairy cattle houses: A scale model study. *Biosystems Engineering*, 2011, 109(2):148-157.
- 54) Petersen, S. O., Andersen, A. J., Eriksen, J., 2012. Effects of cattle slurry acidification on ammonia and methane evolution during storage. *Journal of environmental quality*, 41(1), 88-94.
- 55) Petersen, S.O., Dorno, N., Lindholst, S., Feilberg, A., Eriksen, J., 2013. Emissions of CH<sub>4</sub>, N<sub>2</sub>O, NH<sub>3</sub> and odorants from pig slurry during winter and summer storage. *Nutr. Cycl. Agroecosys.* 95(1), 103-113.
- 56) Petersen, S.O., Højberg, O., Poulsen, M., Schwab, C., Eriksen, J., 2014. Methanogenic community changes, and emissions of methane and other gases, during storage of acidified and untreated pig slurry. *J. Appl. Microbiol.* 117(1), 160-172.
- 57) Portejoie, S., Martinez, J., Guizoui, F., Coste, C., 2003. Effect of covering pig slurry stores on the ammonia emission processes. *Bioresour. Technol.* 87(3), 199-207.
- 58) Pratt, C., Walcroft, A. S., Tate, K. R., Ross, D. J., Roy, R., Reid, M. H., Veiga, P. W., 2012. Biofiltration of

- methane emissions from a dairy farm effluent pond. *Agriculture, ecosystems & environment*, 152, 33-39.
- 59) Rodhe, L. K., Ascue, J., Willén, A., Persson, B. V., Nordberg, Å., 2015. Greenhouse gas emissions from storage and field application of anaerobically digested and non-digested cattle slurry. *Agriculture Ecosystems & Environment*, 199, 358-368.
  - 60) Sakamoto, N., Tani, M., Navarrete, I. A., Koike, M., Umetsu, K., 2008. Covering dairy slurry stores with hydrophobic fertilisers reduces greenhouse gases and other polluting gas emissions. *Australian Journal of Experimental Agriculture*, 48(2), 202-207.
  - 61) Schneider, A. F., Almeida, D. S. D., Yuri, F. M., Zimmermann, O.F., Gerver, M.W., Gewehr, C.E., 2016. Natural zeolites in diet or litter of broilers. *British Poultry Sci* 57, 257-263.
  - 62) Shah, S.B., Grimes, J.L., Oviedo-Rondón, E.O., Westerman, P.W., 2014. Acidifier application rate impacts on ammonia emissions from US roaster chicken houses. *Atmos. Environ.* 92, 576-583.
  - 63) Shin, S. R., Im, S., Mostafa, A., Lee, M. K., Yun, Y. M., Oh, S. E., Kim, D. H., 2019. Effects of pig slurry acidification on methane emissions during storage and subsequent biogas production. *Water research*, 152, 234-240.
  - 64) Sokolov, V. K., VanderZaag, A., Habtewold, J., Dunfield, K., Wagner-Riddle, C., Venkiteswaran, J. J., Gordon, R., 2020. Dairy manure acidification reduces CH<sub>4</sub> emissions over short and long-term. *Environmental technology*, 1-8.
  - 65) Sommer, S. G., Clough, T. J., Balaine, N., Hafner, S. D., Cameron, K. C., 2017. Transformation of organic matter and the emissions of methane and ammonia during storage of liquid manure as affected by acidification. *Journal of environmental quality*, 46(3), 514-521.
  - 66) Sommer, S. G., Clough, T. J., Chadwick, D., Petersen, S. O., 2013. Greenhouse gas emissions from animal manures and technologies for their reduction. *Animal Manure Recycling: Treatment and Management*. Hoboken, New Jersey: John Wiley & Sons, 177-194.
  - 67) Sommer, S. G., Petersen, S. O., Søgaard, H. T., 2000. Greenhouse gas emission from stored livestock slurry. *Journal Environmental Quality*, 29, 744-751.
  - 68) Tao, X., Matsunaka, T., Sawamoto, T., 2008. Dicyandiamide application plus incorporation into soil reduces N<sub>2</sub>O and NH<sub>3</sub> emissions from anaerobically digested cattle slurry. *Aust. J. Exp. Agr.* 48(2), 169-174.
  - 69) Ulens, T., Millet, S., Van Ransbeeck, N., Van Weyenberg, S., Van Langenhove, H., Demeyer, P., 2014. The effect of different pen cleaning techniques and housing systems on indoor concentrations of particulate matter, ammonia and greenhouse gases (CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O). *Livestock Science*, 159, 123-132.
  - 70) Umetsu, K., Kimura, Y., Takahashi, J., Kishimoto, T., Kojima, T., Young, B., 2005. Methane emission from stored dairy manure slurry and slurry after digestion by methane digester. *Animal Science Journal*, 76 (1), 73-79.
  - 71) VanderZaag, A. C., Baldé, H., Crolla, A., Gordon, R. J., Ngwabie, N. M., Wagner-Riddle, C., MacDonald, J. D., 2018. Potential methane emission reductions for two manure treatment technologies. *Environmental technology*, 39(7), 851-858.
  - 72) Wang, K., Huang, D., Ying, H., Luo, H., 2014a. Effects of acidification during storage on emissions of methane, ammonia, and hydrogen sulfide from digested pig slurry. *Biosys. Eng.* 122, 23-30.
  - 73) Wang, Y., Dong, H., Zhu, Z., Li, L., Zhou, T., Jiang, B., Xin, H., 201). CH<sub>4</sub>, NH<sub>3</sub>, N<sub>2</sub>O and NO emissions from stored biogas digester effluent of pig manure at different temperatures. *Agriculture, Ecosystems & Environment*, 217, 1-12.
  - 74) Wang, Y., Dong, H., Zhu, Z., Liu, C., Xin, H., 2014b. Comparison of air emissions from raw liquid pig manure and biogas digester effluent storages. *Transactions of the ASABE*, 57(2), 635-645.
  - 75) Wang, Y., Guo, H., Wang, S., Zhang, J., Zhu, Z., Li, X., Dong, H., 2021. Sulfuric acid modified expanded vermiculite cover for reducing ammonia emissions from animal slurry storage. *Journal of Hazardous*

*Materials*, 403, 123954.

- 76) Wiedemann, S., Phillips, F.A., Naylor, T.A., McGahan, E., Keane, O., Warren, B., Murphy, C., 2016. Nitrous oxide, ammonia and methane from Australian meat chicken houses measured under commercial operating conditions and with mitigation strategies applied. *Anim. Production Sci.* 56, 1404-1417.
- 77) Wood, J. D., VanderZaag, A. C., Wagner-Riddle, C., Smith, E. L., Gordon, R. J., 2014. Gas emissions from liquid dairy manure: complete versus partial storage emptying. *Nutrient Cycling in Agroecosystems*, 99(1-3), 95-105.
- 78) Wu-Haan, W., Powers, W.J., Angel, C.R., Iii, C.E.H., Applegate, T.J., 2007. Effect of an Acidifying Diet Combined with Zeolite and Slight Protein Reduction on Air Emissions from Laying Hens of Different Ages. *Poultry Sci.* 86, 182-190.
- 79) Xue, N., Wang, Q., Wu, C. F., Zhang, L. H., Xie, W. M., 2010. Enhanced removal of NH<sub>3</sub> during composting by a biotrickling filter inoculated with nitrifying bacteria. *Biochem. Eng. J.* 51(1), 86-93.
- 80) Yamulki, S., 2006. Effect of straw addition on nitrous oxide and methane emissions from stored farmyard manures. *Agriculture, Ecosystems & Environment*, 112(2-3), 140-145.
- 81) Yang, J., Wang, C.Q., Cai, Y., Bai, G.C., You, L.Y., Yi, Y.L., Huang, F., Li, X.X., 2015. Life cycle greenhouse gases emission of rice production with pig manure application. *Chinese Journal of Eco-Agriculture* ,23(9), 1131-1141.(in Chinese with English abstract)
- 82) Yang, Y., Sun, Q., Li, N., Liu, C.S., Li, J., Liu, B.S., Zou, G.Y., 2015. Effects of superphosphate addition on NH<sub>3</sub> and greenhouse gas emissions during vegetable waste composting. *Chinese Journal of Applied Ecology* , 26 ,161-167(in Chinese with English abstract)
- 83) Zhou T L, 2017. Characteristics of Gas Emission during Composting of Manure and Dead Pigs and Effect of Biofilter Technology. Chinese Academy of Agricultural Sciences, Beijing.
- 84) Zhu, H. S., Dong, H. M., Zuo, F. Y., Rao, J., 2014. Effect of covering on greenhouse gas emissions from beef cattle solid manure stored at different stack heights. *Transactions of the CSAE*, 30(24), 225-231. (in Chinese with English abstract)
- 85) Zhu, H. S., Zuo, F. Y., Dong, H. M., Luan, D. M., 2017. Effects of sawdust addition and the mixing ratio on ammonia and greenhouse gas emission from stored cattle manure. *Journal of southwest University (Natural Science Edition)*. 39(3), 34-40. (in Chinese with English abstract)